

1 **8**2 **Alberta Oil Sands Development: Risks to Canadian**
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7 The bitumen (oil sand) deposits of northeast Alberta are profound and generally acknowledged
8 as the second largest remaining global reserve of oil. The size of the resource combined with the
9 region's skilled workforce and proximity to the United States make Alberta's oil sands perhaps
10 the most attractive unconventional oil deposit in the world. The vast majority of Alberta's oil
11 sands are yet to be developed due to the high cost of production relative to conventional reserves.
12 The lag in Peak Oil for Alberta's oil sands will create high economic and political pressure to
13 develop the resource to help fill the gap left by declining availability of conventional reserves.
14 Increasing anxiety about the security of Middle East oil is another factor contributing to the
15 increasing availability of United States risk capital to develop this resource (Figure 8-1).

16 The infrastructure, resource demands, and effluents associated with production of
17 Alberta's oil sands could be highly detrimental to wildlife and other environmental values,
18 especially if ecological considerations are marginalized in the rush to develop the resource. In
19 contrast, best management practices and an expanded protected areas network have the potential
20 to offset at least some of the environmental impacts, and the high value of the resource and the
21 region's stability should provide the economic and institutional ingredients needed to achieve

22 high environmental standards. Indeed, access to key markets such as the United States may
23 require ambitious environmental efforts due to the increased ecological literacy of consumers.

24 In this Chapter, we explore the environmental impacts of bitumen development in
25 northeastern Alberta and assess the potential effectiveness of best practices, access management,
26 and protected areas network expansion. Ecological consequences of historical development,
27 including both conventional and bitumen oil extraction, are first chronicled to provide context.
28 We applied the ALCES[®] land use simulation tool (www.alces.ca) to project the ecological
29 impacts of future anticipated development in the region and assess the role of best practices in
30 mitigating risk to ecological indicators. We acknowledge the willingness of the Alberta Land-
31 use Secretariat to provide us with Lower Athabasca Regional Plan (LARP) data to complete
32 these analyses. We further discuss the potential design and function of an expanded protected
33 areas network. The land use simulations and conservation planning analyses that we present
34 were completed to inform the Alberta Land-use Framework (ALUF), a regional land use
35 planning process being coordinated by the Government of Alberta. We conclude the Chapter by
36 describing the Framework and, more generally, the role of proactive planning in balancing the
37 economic and ecological consequences of developing the world's remaining hydrocarbon
38 reserves.

39 **Alberta's Oil Sands Region**

40 The Athabasca Oilsands, the largest of the bitumen deposits in Alberta, are located in the boreal
41 forest natural region. Broad lowland plains and extensive hill systems typify it, with the primary
42 physiographic regions including the Northern Alberta Lowlands, Northern Plains, and
43 Saskatchewan Plains. Prominent topographic features include the Birch Mountains, Stony
44 Mountains, Firebag Hills, and Richardson Sand Hills. A heterogeneous network of deciduous

45 (aspen, birch) and coniferous (white spruce, black spruce, jack pine) forests intermixed with
46 topographically depressed fen and bog complexes, scattered lakes, and main stem rivers and their
47 tributaries characterize this bio-region. The Athabasca River Basin is the primary drainage basin
48 in the region. Gray Luvisols with some Dark Gray Chernozemics soils found in the southern
49 region dominate soils. Organic soils are common within the depressed regions containing fens
50 and bogs. Temperatures are highly variable by season, with average maximum temperatures in
51 the summer of 25° C, and average maximum winter temperatures of -15° C. Average minimum
52 temperatures are 10° C in the summer and -20° C in the winter. Average precipitation (1951-
53 1980) is 450-500 mm/year. Fire is the most important natural perturbation defining forest age,
54 though arthropod outbreaks occur occasionally.

55 The biota of the region reflects the diverse landforms and plant communities of northeast
56 Alberta, including 40 fish species (Nelson and Paetz 1992), five amphibians (Russell and Bauer
57 1993), one reptile (Russell and Bauer 1993), 236 birds (Francis and Lumbis 1979, Semenchuk
58 1992), and 45 mammals (Pattie and Hoffman 1992, Smith 1993). Based on distribution maps
59 (Moss 1983, Vitt et al 1988), conservative estimates indicate a rich diversity of plants, including
60 600 vascular species, 17 ferns, 104 mosses, 13 liverworts, and 118 lichen species.

61 Although challenging to quantify, natural capital, such as biodiversity, provides
62 important ecosystem services, such as water resources, sequestered carbon, and climate
63 regulation. The non-market ecological goods and services (based on 17 elements) for the
64 MacKenzie River Basin is estimated at \$2,631/ha/year (Anielski and Wilson 2007). Concerns
65 expressed by the environmental community emphasize the potential long-term harm that could
66 occur to the region's EGS by maximizing the short-term gain in GDP associated with extraction
67 and processing of bitumen.

68 The findings of this report address the LARP region of northeast Alberta, an area that
69 directly overlaps the Athabasca and Cold Lake bitumen deposits. The LARP region represents
70 the first region to develop a strategic management plan under the auspices of the Alberta Land
71 Stewardship Act passed in 2009.

72 Although this region has supported First Nation communities for thousands of years, and
73 trapping for hundreds of years, it is only during recent decades that large-scale industrial
74 development has emerged. The two dominant land uses, in terms of area affected, are the
75 forestry and energy sectors. Both land uses have grown exponentially in harvest and extraction
76 volumes during the past few decades (Figures 8-2 and 8-3). In addition to Alberta-Pacific Forest
77 Industries, Inc. (Al-Pac), which holds a large hardwood-dominated Forest Management
78 Agreement (FMA) that overlaps spatially with LARP, there exist a few smaller softwood
79 allocations for quota holders. The Annual Allowable Cut (AAC) of Al-Pac is currently ~3.5
80 million m³/year and about 2.7 million m³/year of that full volume is harvested from the LARP
81 study area.

82 It is difficult to quantify total volumes of bitumen “in-place”, but most sources provide
83 estimates in the range of ~1,804 billion barrels (287 billion m³) (ERCB 2010). Of this volume,
84 existing technologies that are economically viable could remove ~170 billion barrels (27 billion
85 m³) (ERCB 2010). This volume is likely to increase as newer technologies emerge. The volume
86 of bitumen removed to date (~6.9 billion barrels, 1.1 billion m³) represents less than 4% of the
87 recoverable volume. Given assumptions of approved and projected new bitumen projects, the
88 current annual production levels of 1.49 M bpd (0.24 M m³/year) is expected to increase to 3.2 M
89 bpd (0.51 M m³/year) by 2019 (ERCB 2010).

90 **Impacts of Future Oil Sands Development and Opportunities for**

91 **Mitigation**

92 Development of unconventional deposits such as oil sands is expected to accelerate in the future,
93 in part to fill the gap in oil supply created by diminishing conventional reserves. Investment and
94 production projections for Alberta's oil sands suggest a rapidly increasing rate of development
95 (Figure 8-4). Concomitant with expanded oil sands production is increased environmental
96 impacts due to growth in landscape disturbance, water and energy consumption, and emissions.
97 Accelerated growth of the industry has raised concerns regarding reduced terrestrial and aquatic
98 wildlife habitats due to its industrial footprint (Schneider and Dyer 2006) and reduced river flow
99 (Schindler et al. 2007), water contamination (Holroyd and Simiertsch 2009), wildlife mortality at
100 tailings ponds (Wells et al. 2008), loss of biological carbon storage (Lee and Cheng 2009),
101 increased CO₂ emissions (Raynolds et al. 2006), and acid rain (Prebble et al. 2009).

102 **Exploring the Future Consequences of Development**

103 We evaluated the potential future effects of expected development trajectories using the ALCES
104 land use simulation tool. ALCES generates projections for a wider range of environmental and
105 socioeconomic indicators under alternative assumptions about land-use policies and ecological
106 processes. Here we present results for the following subset of the wide array of biodiversity,
107 water, land, and economic indicators that were assessed in simulations to inform a land use
108 planning process being led by the Government of Alberta (ALCES Group 2009): anthropogenic
109 footprint, forest age, water consumption, woodland caribou, an index of fish community
110 integrity, and carbon dioxide and nitrous oxide emissions. Simulations explored the response of
111 these indicators to development trajectories 100 years into the future.

112 ALCES is well suited to assess strategic-level implications of development trajectories in
113 the region due to its capacity to simulate the cumulative effects of the major types of land use
114 (hydrocarbon extraction, forestry, agriculture, residential) and natural processes (fire, insect
115 outbreaks, and meteorology). Using an annual time-step, the model applies natural disturbances
116 and successional trajectories as well as anthropogenic footprints (well sites, mines, seismic lines,
117 pipelines, roads, cut blocks, settlements, and farmland) associated with simulated resource
118 production rates that have user-defined life spans. The first-order impacts tracked by ALCES are
119 on resource supply and landscape composition, which are then translated into a variety of
120 indicator variables using equations.

121 To provide a starting point for the simulations, the current composition of the 93,000-km²
122 study area was quantified using inventories of vegetation and anthropogenic features.
123 Extractions of two types of hydrocarbons were included in simulations: mineable bitumen that is
124 extracted using open-pit mines and in-situ bitumen that is extracted using technologies such as
125 steam-assisted gravity drainage (SAGD) and cyclic steam stimulation (CSS). Bitumen
126 development followed a pattern of increasing extraction to peak production followed by a
127 gradual decline, as per the Hubbert-Naill model (Naill 1973). Reserve sizes and production rates
128 were based on government projections (Figure 8-4), as were intensities, such as disturbance per
129 unit of resource production, for seismic lines, well pads, and pipelines (Bob Nichol, consultant to
130 Alberta Energy, personal communication). ALCES also attempted to maintain the annual
131 allowable wood harvest allocated by the government to forestry companies in the region (2.75
132 million m³/yr). Available merchantable timber was based on growth and yield curves, and
133 constrained by harvest eligibility of forest types and age classes and by deletions from the active
134 forestland base to account for inaccessible stands and protected areas. Forests were also

135 disturbed by fire in order to assess the cumulative effect of natural and anthropogenic
136 disturbances. A 1.25% annual burn rate across forest types and seral stages was simulated based
137 on fire research for the region (Armstrong 1999). Although the regional fire regime is stochastic,
138 fire was simulated as deterministic so that random fluctuations in annual burn rate did not
139 obscure the effects of alternative land-use strategies. Disturbed stands were assumed to
140 regenerate to their pre-disturbance forest type.

141 Other land uses in the region include agriculture and road network expansion. Future
142 agricultural activity was simulated based on a government-endorsed projection that current
143 cropland area (240,554 ha) will expand by 350 ha/year over the next 50 years and 700 ha/year
144 thereafter. Highways grew at a rate of 35 km/year, while secondary roads grew by an average of
145 31 km/year. In addition, roads required to access timber were assumed to cover 3% of cut blocks
146 and 0.5 km of access road was assumed associated with each well pad. Highways were
147 simulated as permanent features, whereas secondary, in block, and well pad access roads had life
148 spans of 25, 10, and 30 years, respectively. Culverts were applied to secondary, in block, and
149 well pad access roads to account for intersections between road and stream networks. Culverts
150 can be detrimental to fish habitat if they “hang”, which refers to the tendency of the downstream
151 end of a culvert to become suspended above the stream because of scouring by outflow.

152 Hanging culverts can prevent fish passage, thereby fragmenting fish habitat and potentially
153 decreasing access to fish spawning and rearing areas (Park et al. 2008). The abundance of
154 hanging culverts in simulations was estimated based on the density and age-class distribution of
155 culverts.

156 Simulations explored two strategies for mitigating the potential impacts of development
157 in the region: best practices and access management. Best practices refer to the best available

158 strategies, within technological and economic constraints, for limiting environmental degradation
159 per unit of resource production. Access management refers to reducing motorized public
160 movement along a suite of industrial and natural features, including roads, trails, seismic lines,
161 pipelines, watercourses and lakes. Increased public access facilitated by an expanding industrial
162 footprint can increase angling and hunting, activities that may contribute to fish and wildlife
163 declines in the region (Sullivan 2003). Best practices and access management were simulated at
164 moderate and high levels to consider the implications of realistic and optimistic assumptions
165 regarding the effectiveness of these strategies. Seven scenarios defined by various levels of best
166 practices and access management were explored: business as usual (BAU: no best practices or
167 access management); moderate best practices combined with each level of access management
168 (none, moderate, and high for a total of three simulations); and high best practices combined
169 with each level of access management (for a total of three simulations). The set of seven
170 simulations allowed us to explore the incremental benefits of implementing best practices and,
171 subsequently, access management.

172 The suite of best practices assessed in simulations was identified through consultations
173 with experts from relevant government departments and industry. Included were strategies for
174 minimizing the size and duration of footprint, old forest protection, water conservation, and
175 emissions reduction (Table 8-1). The effect of access management scenarios on fish and wildlife
176 was based on the outcomes of workshops held with government biologists and wildlife managers
177 (Sullivan 2009). Moderate and high access scenarios were defined as 50% and 75% reductions,
178 respectively, in public motorized use through road restrictions, designated trails, and motor
179 restrictions on lakes. Under the BAU scenario, access management was defined as being
180 generally unregulated leading to extensive human access and heavy use (angling) at most lakes.

181 Woodland caribou and an index of native fish integrity were evaluated to consider
182 impacts to terrestrial and aquatic wildlife. Woodland caribou is a focal species for the region due
183 to its threatened status and sensitivity to both forest age and industrial footprint density. The
184 effect of simulated landscape transformations on woodland caribou was assessed using a model
185 that relates woodland caribou finite rate of increase (λ) to anthropogenic edge density
186 and forest younger than 30 years (Boutin and Arienti 2008). Based on caribou data from
187 Alberta, the model estimated the relationship between caribou finite rate of increase and two
188 attributes of landscape composition (linear features and forest younger than 30 years) by the
189 following formula:

190 Finite rate of increase = $(1.0184 - 0.0234) \times (\text{linear feature density} - 0.0021) \times \text{percent of forest}$
191 younger than 30 years

192 A sustained λ value less than 1 implies eventual extirpation of caribou from the region.
193 Moderate and high access management was assumed to increase λ by 0.0075 and 0.015,
194 respectively, based on expert opinion (Sullivan 2009). Impacts of development on the fish
195 community were assessed using the Index of Native Fish Integrity (INFI), a measure that
196 conveys changes in abundance and composition of fish species with a value ranging from 1.0
197 (undisturbed community) to 0 (highly disturbed). The relationships between INFI and human
198 access, stream network fragmentation, and water use were based on a workshop held with
199 regional fishery experts (Lagimodiere and Eaton 2009).

200 In addition to evaluating indicators at the regional scale, subregional performance was
201 assessed by developing maps of simulated landscape composition and indicator status midway
202 into the simulation period (year 60). Maps were produced using ALCES Mapper, a companion-

203 mapping tool for ALCES that creates spatial representations that are consistent with ALCES
204 outputs. Rather than apply industrial development uniformly across the study area, the location
205 of development was influenced by existing industrial infrastructure, the spatial distribution of
206 resources such as bitumen and timber, and land use zones such as protected areas that were
207 excluded from development. Maps of caribou and INFI status were created based on spatial
208 representations of simulated landscape composition and access management. Access
209 management varied spatially, with reduced access management in agricultural and settled areas,
210 and the mineable bitumen zone. When mapping INFI, status was expressed in terms of risk.
211 Risk was assessed based on percent departure from the indicator's estimated Range of Natural
212 Variability (RNV) as estimated from 100 simulations that included stochastic meteorology, but
213 excluded industrial development. The risk levels, based on criteria developed by the
214 International Union for Conservation of Nature (IUCN), were as follows: stable (less than 10%
215 from RNV); low risk (10–50% from RNV); moderate risk (50–70% from RNV); and high risk
216 (greater than 70% from RNV).

217 **Simulation Results**

218 Anthropogenic footprint increased rapidly from approximately 7–10% of the landscape during
219 the first 25 years of the simulation as development outpaced reclamation. By the end of the
220 simulation, the footprint covered over 11% of the landscape and was focused in the southern and
221 central portions of the study area where bitumen and timber reserves are prevalent. Best
222 practices have the potential to achieve large reductions in the energy sector footprint through
223 accelerated reclamation and reduced footprint intensity. When best practices were applied,
224 simulated peak density of anthropogenic edge was approximately half of that projected in the
225 BAU scenario (Figures 8-5 and 8-6). Another large landscape alteration apparent from the

226 simulations was reduced abundance of older forest, a consequence of forestry activity in the
227 region (Figure 8-7). In contrast to the projected changes in landscape composition, the simulated
228 impact of land use on river flow was relatively minor. Water extraction by industry and
229 settlements peaked at almost 8% of main stem river flow in February when riverine systems in
230 the region experience their lowest flow and are therefore most susceptible to water removal.

231 Caribou finite rate of increase (λ) associated with the current landscape was estimated to
232 equal 0.95, suggesting that the existing landscape is not capable of supporting a viable caribou
233 population. This conclusion is supported by an assessment of management options for the
234 region's caribou herds that determined that functional habitat was insufficient to maintain
235 caribou in the region beyond two to four decades (Athabasca Landscape Team 2009). In the
236 absence of best practices, λ was simulated to decline over the next quarter century and then to
237 slowly recover (Figure 8-8). The pattern mirrors the simulated trend in anthropogenic edge
238 density (Figure 8-5), demonstrating the influence of the industrial footprint on the species. It is
239 thought that the anthropogenic footprint is detrimental to caribou because it promotes invasion
240 by moose, deer, and wolves, ultimately leading to levels of predation that the caribou population
241 cannot sustain (James et al. 2004). Woodland caribou are susceptible to predation due to their
242 low reproductive rate and inability to escape predators, such as wolves. By minimizing and
243 reclaiming industrial footprint, implementation of best practices decreased the maximum
244 anthropogenic edge density by over 50% and, as a result, improved simulated λ relative to the
245 BAU simulation (Figure 8-8). Access management had the effect of further improving λ ,
246 although the improvement was small because human access (hunting or vehicle collisions) is
247 thought to be only a minor contributor to caribou mortality. Importantly, λ remained below 1
248 across all simulations suggesting that caribou are likely to be extirpated from the region despite

249 aggressive mitigation strategies, with the exception of caribou herds located in the northern
250 portion of the study area where bitumen deposits do not exist (Figure 8-9).

251 As with caribou, the current landscape was estimated to support a degraded fish
252 community relative to natural conditions. This assessment is consistent with fish surveys and
253 research in northern Alberta that have demonstrated wide-spread declines in recreational
254 fisheries, primarily due to unsustainable rates of angling (Post et al. 2002). The INFI was
255 simulated to continue to decline with future development (Figure 8-10), approaching a value of 0
256 which is indicative of fish communities dominated by small fish such as suckers, minnows, and
257 chub and almost devoid of sport fish such as walleye, pike, and Arctic grayling (Lagimodiere
258 and Eaton 2009). Contributing to the decline was fragmentation of the stream network as
259 hanging culverts became more prevalent due to increased road density. Relative to BAU, best
260 practices reduced stream network fragmentation through improved culvert maintenance and
261 reduced road density. As a result, the decline in INFI was not as sharp in the best practice
262 simulation. Even with best practices, however, the fish community was well below natural
263 conditions, largely due to angling pressure that increased as the region's population grew.

264 The failure of fishing regulations to sustain northern Alberta fisheries is due to both the
265 region's cold climate and the high angling pressure. Cool temperatures result in a low
266 productivity fishery that requires more than 10 years to produce large fish. Angling pressure is
267 two orders of magnitude higher than fisheries in other Canadian provinces such as
268 Saskatchewan, Manitoba, and Ontario due to the rapidly growing human population and the
269 relatively small number of lakes (Sullivan 2003). When access management was simulated,
270 INFI improved dramatically due to a drop in angling pressure (Figure 8-10). The dramatic
271 response of INFI to human access is consistent with boreal fishery research (Post et al. 2002,

272 Sullivan 2003) and emphasizes the susceptibility of the region's fisheries to angling pressure.
273 Simulated future status of INFI differed substantially across watersheds, reflecting differences in
274 fragmentation and, especially, angler access (Figure 8-11). Watersheds with low human
275 population density, such as the far north, or high access management were simulated to support
276 more intact fish communities.

277 In addition to disturbing landscapes, oil sands production emits a variety of pollutants.
278 Oil sands CO₂ emissions intensity is greater than that of conventional oil, primarily because
279 more energy is required during refining. The high emission intensity combined with rapidly
280 escalating production makes the oil sands a significant contributor to greenhouse gas (GHG)
281 emissions. Today's emissions of approximately 40 megatonnes CO₂-equivalent (Mt CO₂-e)
282 accounts for 5% of Canada's GHG emissions of 747 Mt CO₂-e in 2007 (Environment Canada
283 2007). Due to continued growth in oil sands production, CO₂ emissions were simulated to more
284 than triple within the next 25 years, and then gradually decline as production diminishes. Peak
285 simulated emissions of 130 Mt CO₂-e accounts for almost one quarter of Canada's Kyoto target
286 of 558.4 Mt CO₂-e, emphasizing that controlling oil sands GHG emissions represents a
287 substantial challenge. The government of Alberta estimates that emissions intensity can be
288 reduced by almost half using strategies that are within technological and economic constraints.
289 Implementing these practices reduced peak GHG emissions to just over 60 Mt CO₂-e. While
290 this is an obvious improvement compared to existing practices, it still represents a 50% increase
291 in GHG emissions relative to today. More ambitious emissions intensity efforts (Raynolds et al.
292 2006) or reduced production or both may therefore be required if oil sands production is to be
293 within the bounds of an increasingly carbon-constrained world.

294 Other emissions of concern are potential acid inputs such as sulphur dioxide and nitrous
295 oxides due to their contribution to acid deposition. The study area contains waters that are
296 susceptible to acidification due to the low alkalinity of some lakes located in the Canadian Shield
297 and upland northern areas (Trew 1995). Acidification can be very detrimental to aquatic
298 ecosystems, in severe cases leading to large reductions in species diversity (Schindler et al.
299 1985). Nitrous oxide and sulphur dioxide emissions were projected to almost double in the BAU
300 simulation. At their peak, projected nitrous oxide emissions exceeded the 86,000 tonnes/year
301 preliminary regional threshold for nitrous oxides by approximately 100%. Although best
302 practices were able to reduce substantially simulated nitrous oxide emissions, the emission
303 threshold was still surpassed by 30,000 tonnes/year. It would therefore appear that lakes in the
304 region might be at risk of acidification due to projected growth in oil sands production.

305 Bitumen extraction in the region is likely to increase rapidly, in part to fill the production
306 gap caused by declining conventional reserves. Landscape disturbance, human access, and
307 emissions associated with projected rates of production would contribute to continued decline in
308 regional fish and wildlife populations and increased greenhouse gas emissions. Application of
309 best practices has the potential to reduce substantially disturbance and emission intensities.
310 Results from simulations demonstrate, however, that the reduction was insufficient to avoid
311 declines in wildlife due to the rapid escalation of production. Access management was highly
312 successful at improving INFI, emphasising the need to limit angling pressure if the region's fish
313 populations are to recover. Access management was modestly successful for improving the
314 status of black bear, moose, and wolverine, but did not contribute meaningfully to maintenance
315 of regional caribou populations. In the next section, we consider the potential role that
316 conservation areas could play in mitigating regional impacts to wildlife.

317 **The Role of Conservation Areas**

318 Even in the presence of aggressive efforts to minimize impacts, disturbance caused by bitumen
319 development in the region will be substantial. Allocating a portion of the landscape to
320 conservation is a potential mechanism for offsetting at least some of the impacts on wildlife.
321 Protected areas can contribute to regional conservation goals by representing ecological
322 community types, preserving rare features, maintaining habitat for species negatively affected by
323 development, and providing examples of natural ecosystem function. Existing protected areas
324 networks are typically inadequate to achieve these goals (see also Czech 2005, Dietz and Czech
325 2005). Twelve percent of the world's landscapes are protected (Chape et al. 2003), yet
326 representation of all species and ecosystem types in a region typically requires substantially
327 higher levels of protection, typically 25-75% (Noss and Cooperrider 1994). Although protected
328 areas presently cover only seven percent of the LARP region, the government has expressed its
329 intent to increase protection to at least 20% of the region. This is consistent with the region's
330 multi-stakeholder Cumulative Environmental Management Association that recommended 20-
331 40% of the region be set aside for conservation (SEWG 2008).

332 To assess the potential contribution of increased protection, we considered design
333 requirements of conservation areas to improve on the goals of representation and ecological
334 benchmarks. Maintaining representative examples of ecological community types is a coarse-
335 filter approach to conservation planning that is based on the assumption that conserving the
336 range of native community types will also conserve the majority of species. Although this
337 assumption is largely untested, representation remains the only practical strategy for identifying
338 conservation areas that potentially represent a region's biodiversity. Ecological benchmarking
339 refers to the concept of maintaining examples of natural ecosystems to provide controls for

340 management experiments needed to develop knowledge for sustainable management
341 (Schmiegelow et al. 2006).

342 While beneficial for conservation goals, ecosystem protection reduces availability of
343 resources for economic growth. Protection can therefore come into conflict with socioeconomic
344 goals. We evaluated the economic implications of an expanded conservation areas network by
345 assessing the degree to which bitumen reserves would be made inaccessible. Efforts were made
346 to balance economic and ecological objectives by minimizing the presence of bitumen reserves
347 within protected areas. This is consistent with the triad land use zoning strategy that seeks to
348 overcome conflict between ecological and economic objectives by dividing a region's land base
349 into three categories: mixed-use (also referred to as extensive), protection, and intensive-use.
350 High resource production generated by the intensive-use component of the landscape makes up
351 for the drop in production caused by protection and ecosystem-based management practices.
352 The triad was developed as a strategy to maintain timber supply through forest plantations while
353 increasing the use of sustainable forestry practices and protected areas on the remainder of the
354 landscape (Hunter 1990). The zoning strategy is equally applicable to the production of other
355 commodities from the forestland base including fuel (Hunter 2002), and is potentially well suited
356 to the LARP region due to the intensive nature of bitumen development and the aggregated
357 distribution of bitumen deposits (Figure 8-12). The triad has been proposed as a strategy for
358 balancing the region's ecological and economic objectives (SEWG 2008).

359 We adopted a 30% representation goal of each the study area's eight natural subregions
360 and fourteen land cover types. Natural subregion is a coarse level of ecosystem classification
361 characterized by vegetation, climate, elevation, and latitudinal or physiographic differences
362 (Natural Regions Committee 2006). Land cover reflects the finer scale mosaic of forest,

363 wetland, and aquatic ecosystem types that occur across the study area. Exploratory analyses
364 determined that using natural subregions and land cover types as representation attributes struck
365 a suitable balance between ecological detail and economic cost. Thirty percent was selected as
366 the representation target because it is the mid-point of the 20-40% protected area goal
367 recommended for the region by the Cumulative Environmental Management Association
368 (SEWG 2008). A conservation areas network satisfying the representation goals was designed
369 using MARXAN, software that applies optimization to select a conservation area network that
370 achieves representation targets at the least cost (Ball and Possingham 2000, Possingham et al.
371 2000). The analysis used 10 km² cells (townships) as building blocks for the conservation areas
372 network, and forced MARXAN to include existing protected areas and exclude areas that are
373 heavily disturbed (greater than 10% converted). The conservation area design with the lowest
374 cost identified by MARXAN (Figure 8-13) achieved the 30% representation goals at a cost of
375 excluding 3.7% of bitumen volume from development. It is therefore apparent that protecting a
376 substantial and representative portion of the planning region to safeguard regional biodiversity
377 from future development can be achieved at low economic cost. The exception is the southern-
378 most natural subregion, which is already too disturbed by agriculture to make the 30%
379 representation goal feasible.

380 To identify potential ecological benchmarks in the planning region we collaborated with
381 the Canadian BEACONS Project (BEACONS), a research initiative led by Dr. Fiona
382 Schmiegelow of the University of Alberta that is focused on developing conservation area
383 strategies that provide ecological benchmarking capacity to facilitate adaptive management of
384 boreal ecosystems. To function as an ecological benchmark, a conservation area should be
385 representative of the natural conditions of the region, free from human activities that interfere

386 with ecological processes, and large enough to capture ecological processes and maintain all
387 species in the presence of the fire regime (Canadian BEACONS Project 2008). Of these
388 characteristics, size is typically the most constraining requirement and the majority of existing
389 conservation areas are too small to act as ecological benchmarks. To act as an ecological
390 benchmark, a conservation area should be larger than the maximum expected natural disturbance
391 to provide internal recolonization sources. The planning region's largest conservation area
392 (200,000 ha) is substantially smaller than the maximum fire size (500,000 ha), underscoring the
393 need to improve benchmarking capacity.

394 To consider strategies for improving the benchmarking capacity of the conservation areas
395 network, BEACONS identified candidate benchmarks in the planning region using the
396 Benchmark Builder (Version 15). The Benchmark Builder is software designed to automate the
397 design of candidate ecological benchmarks in Canada's boreal forest. When designing candidate
398 benchmarks, the Benchmark Builder considered hydrologic connectivity, landscape intactness,
399 size, and biophysical representation. Hydrologic connectivity and landscape intactness were
400 assessed to identify areas with intact aquatic and terrestrial ecosystems whereas size was
401 assessed to identify areas that were larger than the maximum fire size. Candidate ecological
402 benchmarks identified by the Benchmark Builder were then compared in terms of overlap with
403 bitumen reserves and representation of natural subregions in order to consider compatibility with
404 other conservation area objectives, for example, minimizing cost and ecological representation.
405 The analysis determined that the option best able to provide benchmarking capacity at minimum
406 cost while also contributing to natural subregion goals was creation of two nearby protected
407 areas that spanned a total of 1,156,518 ha (10.9% of the region) and contained 2.0% of the
408 region's bitumen. As was the case with representation goals, we therefore conclude that

409 expanding the conservation areas network to improve ecological benchmarking capacity should
410 not be cost-prohibitive.

411 To summarize, we were able to identify options for expanding the conservation areas
412 network that would protect 30% of each natural subregion and land cover type and establish
413 ecological benchmarks, while excluding only a small portion of the region's bitumen from
414 extraction. We conclude that increasing the conservation areas network is a cost-effective
415 strategy for safeguarding regional biodiversity from increased disturbance that can be expected
416 with accelerated bitumen extraction. The viability of a large protected areas network in the
417 presence of continued bitumen extraction is due to the aggregated distribution of bitumen that
418 permits the establishment of large protected areas that have minimal overlap with bitumen
419 reserves. It is important to note that this compatibility would not hold if the conservation areas
420 network was designed to maintain viable caribou herds. Caribou herds in the region overlap with
421 bitumen deposits and, as a result, are at risk to future development. Maintaining caribou in the
422 region beyond the next couple of decades requires aggressive actions including: the
423 establishment of large (thousands of square kilometres) caribou conservation areas where the rate
424 of development (including bitumen extraction) is strictly controlled; rapid reclamation of
425 existing footprint; and a long-term wolf control program to limit predation (Athabasca
426 Landscape Team 2009). Other ecological impacts will similarly not be addressed through
427 protection alone, such as emissions that contribute to greenhouse gas concentrations and acid
428 deposition. It is therefore apparent that some impacts on expanded bitumen extraction cannot be
429 avoided through management practices or the establishment of cost-effective conservation areas.
430 These unavoidable tradeoffs between ecological integrity and economic activities represent the

431 greatest challenge facing land use planning in the region. Required is a societal decision
432 regarding the appropriate balance between economic growth and ecological sustainability.

433 **Planning Land Use to Balance Hydrocarbon Development and** 434 **Wildlife Conservation**

435 The oil sands and their development trajectory remain a controversial issue at local, regional,
436 provincial, national, and international scales. Much has been written (see Nikiforuk 2008 and
437 Kelly 2009 for an overview) about the potential for Alberta's bitumen-based hydrocarbons to
438 contribute to the North American fuel stock, to sustain Alberta's and Canada's economy, to
439 reduce North American dependency on Middle East oil, and its environmental footprint and
440 relatively "dirty" image in comparison to other energy alternatives.

441 Given that the majority of Alberta's bitumen is currently destined to a UNITED STATES
442 market, UNITED STATES policies (and related market demand and pricing) will weigh heavily
443 on the pace and tempo of bitumen production in Alberta. Recent policy initiatives emerging in
444 the U.S., such as California's Low Carbon Fuel Standard (LCFS), that requires an assessment of
445 the full life cycle carbon intensity of alternative fuel stocks, may influence the attractiveness of
446 Alberta's oil sands to the United States from both an investment and commodity importation
447 perspective. Market demand for future bitumen is further obscured by the emergence of offshore
448 markets, such as the recently announced Enbridge Northern Gateway Project that would move
449 Alberta's bitumen by pipeline to coastal British Columbia, and subsequently by super-tankers to
450 Asian markets, such as China. As such, a potential loss or reduction in United States market
451 demand for bitumen would certainly have a large impact on Alberta's oil sand play, but this
452 theoretical reduction might be offset by new offshore markets.

453 From a land use perspective, Alberta's supercharged economy has spawned a host of
454 related issues and challenges at all spatial scales. A province that only a few decades back was
455 perceived as vast with few people is now viewed as small and crowded with conflicting land use
456 objectives and disgruntled land use players.

457 Against this backdrop, the Alberta government has embarked on an ambitious initiative
458 of regional planning called the Alberta Land-use Framework (ALUF). This integrated planning
459 process, mandated by the recently enacted Alberta Land Stewardship Act, has numerous
460 objectives that focus on creating a strategic vision of land use trajectories that balance economic,
461 social, and environmental objectives in Alberta. This new legislation is omnibus in nature, and
462 effectively trumps other Acts that may be in conflict with its intent. Rather than tackling the
463 entire province as one geographic unit, the ALUF has stratified the province into seven regions
464 that are broadly based on regional watersheds and municipal boundaries. The initial stratum
465 examined by the ALUF is the Lower Athabasca Regional Plan, a region that contains the
466 majority of bitumen deposits in Alberta.

467 Guided by goal-posting directives from Alberta's Cabinet, each ALUF region completes
468 its work using a Regional Advisory Council (about a dozen individuals representing major
469 stakeholder groups) and a Regional Planning Team comprised of government employees
470 providing disciplinary expertise. This work is overseen by the Land-use Secretariat. As their
471 title suggests, the primary goal of the Lower Athabasca Regional Plan is to construct a plan that
472 can be submitted to Cabinet for consideration and, hopefully, approval and implementation. The
473 primary objectives of the plan include (Morris Seiferling, Director of Alberta Land-use
474 Framework, personal communication):

- 475 • A strategic level plan for land and natural resource use on public and private lands that
- 476 defines regional outcomes for economic, environmental and social indicators
- 477 • An alignment of provincial strategies and policies at the regional level
- 478 • A determination of specific trade-offs and appropriate land and natural resource
- 479 management for each region
- 480 • A quantitative description of the cumulative effects of the combined land use trajectories
- 481 and the identification of specific thresholds and targets for key indicators

482 The wording of the Alberta Land Stewardship Act and the ALUF indicate a commitment
483 of the Government of Alberta to adopt an integrated approach to resource stewardship.
484 Examples of integrated approaches include integrated resource management (IRM), integrated
485 environmental management, integrated catchment management, watershed management,
486 bioregional planning, and integrated landscape management. These various monikers all share a
487 common approach characterized by a proactive, holistic, systems-based, and integrated approach
488 to environmental problems.

489 While numerous definitions of integrated resource management exist, the description
490 offered by Cairns and Crawford (1991) is directly relevant to the goals of the ALUF:

491 *“Coordinated control, direction or influence of all human activities in a defined environmental*
492 *system to achieve and balance the broadest possible range of short- and long-term objectives.”*

493 For the ALUF to achieve a reasonable level of success, four essential characteristics that
494 distinguish proper Integrated Resource Management (IRM) from “reactionary” management will
495 need to be respected. These elements are drawn from Wikipedia’s description of IRM prepared
496 by Carlson and Stelfox (<http://www.eolss.net/>).

497 1. **Inclusive**. IRM considers the broad spectrum of ecological, social, political, and economic
498 factors and large spatial and temporal scales that define environmental issues. In contrast to
499 mono-disciplinary management approaches such as sustained yield, IRM demands a
500 multidisciplinary approach that engages diverse perspectives and skill sets. Decision-making is a
501 collaborative process involving the public. There exists an explicit recognition that empirical
502 science alone cannot lead to a solution, but rather that a society informed by science can better
503 arrive at optimal land use trajectories.

504 2. **Interconnective**. IRM evaluates how different components of ecological and human systems
505 interact. This system dynamics approach recognizes that ecosystems are complex systems with
506 emergent properties that cannot be ascertained through reductionism and that, as a result,
507 environmental problems cannot be solved by compartmentalization.

508 3. **Goal-oriented**. IRM is goal-oriented and proactively plans for a desired state. The goals are
509 typically broad and defined through a collaborative process involving diverse stakeholders. The
510 goal-setting process therefore not only fosters a proactive perspective, but also inclusivity and
511 broad ownership in planning outcomes.

512 4. **Strategic**. Goal setting also focuses attention on key elements of the system of concern. This
513 focus is needed to address strategically environmental issues amongst the complexity and
514 uncertainty of environmental systems. IRM's strategic approach is adaptive and intentionally
515 seeks to improve knowledge of the ecological and social effects of land use. At the same time,
516 IRM is precautionary to limit the risk of unanticipated and undesirable impacts.

517 Given the early stages of the ALUF, it is premature to comment on how well it will serve
518 Albertans in defining and implementing sustainable land use trajectories for the seven regions.

519 The Government of Alberta should be applauded for launching this initiative, but stakeholders
520 (public, government, industry) have yet to learn the full extent of tough decisions that attend a
521 well-crafted integrated land use plan.

522 It is clear, however, that the Government's highest priority for the LARP is the extraction
523 of bitumen. Work completed to date by LARP confirms that historical bitumen production, in
524 concert with forestry and other land uses, have significantly compromised the performance of
525 selected ecological indicators, and that increased levels of bitumen production will lead to further
526 reductions in performance. Some of this risk to ecological indicators can be mitigated through
527 the adoption of industrial best practices, the implementation of access management policies, and
528 the designation of an expanded network of protected areas. The balance between bitumen
529 production and resource revenues, on the one hand, and the integrity of ecosystems on the other,
530 has not yet been determined but is the topic of much dialogue within LARP.

531 **Literature Cited**

532 ALCES Group. 2009. Lower Athabasca Regional Plan ACLES III Scenario Modeling Scenario
533 Package 2 (Best Practices). Unpublished report prepared for the Alberta Government
534 Land-Use Secretariat.

535 Anielski, M., and S. Wilson. 2007. The real wealth of the Mackenzie Region. Canadian Boreal
536 Initiative, Ottawa, Ontario, Canada.

537 Armstrong, G. W. 1999. A stochastic characterisation of the natural disturbance regime of the
538 boreal mixedwood forest with implications for sustainable forest management. Canadian
539 Journal of Forest Research 29:424–433.

540 Athabasca Landscape Team. 2009. Athabasca caribou landscape management options report.
541 Unpublished report submitted to Alberta Caribou Committee, Edmonton, Alberta,

- 542 Canada. <<http://www.albertacariboucommittee.ca/PDF/Athabasca-Caribou.pdf>>.
543 Accessed 26 Jul 2010.
- 544 Ball, I. R., and H. P. Possingham. 2000. MARXAN (V1.8.2): Marine reserve design using
545 spatially explicit annealing, a manual.
- 546 Boutin, S., and C. Arienti. 2008. Boreal Caribou Committee (BCC) equation reanalysis, final
547 Report – November 26, 2008. Unpublished report prepared for the Alberta Caribou
548 Committee Research and Monitoring Subcommittee by University of Alberta Integrated
549 Land Management Laboratory, Edmonton, Alberta, Canada.
- 550 Canadian BEACONS Project. 2008. Creating a scientific foundation for conservation planning
551 in boreal Canada. University of Alberta and Canadian Boreal Initiative.
552 <<http://www.beaconsproject.ca/resources.htm>>. Accessed 13 Oct 2009).
- 553 Cairns, J., Jr., and Todd V. Crawford. 1991. Integrated environmental management. Lewis
554 Publishers, Incorporated, Chelsea, Michigan, USA.
- 555 Chape, S., L. Fish, P. Fox, and M. Spalding. 2003. United Nations list of protected areas.
556 IUCN/UNEP, Gland, Switzerland/Cambridge, United Kingdom.
- 557 Czech, B. 2005. The capacity of the National Wildlife Refuge System to conserve threatened and
558 endangered animal species in the United States. *Conservation Biology* 19(4):1246-1253.
- 559 Dietz, R., and B. Czech. 2005. Conservation deficits for the continental United States: an
560 ecosystem gap analysis. *Conservation Biology* 19(5):1478-1487.
- 561 Energy Resources Conservation Board [ERCB]. 2010. Alberta's energy reserves 2009 and
562 supply/demand outlook 2010-2019. ST98-2010. Energy Resources Conservation Board,
563 Calgary, Alberta, Canada. <http://www.ercb.ca/docs/products/STs/st98_current.pdf>.
564 Accessed 24 Aug 2010.

- 565 Environment Canada. 2007. Canada's 2007 greenhouse gas inventory.
566 http://www.ec.gc.ca/pdb/ghg/inventory_report/2007/som-sum_eng.cfm#s2. Accessed
567 13 Oct 2009).
- 568 Francis, J., and K. Lumbis. 1979. Habitat relationships and management of terrestrial birds in
569 northeastern Alberta. AOSERP (Alberta Oil Sands Environmental Research Program)
570 Report No. 78. Prepared for AOSERP by Canadian Wildlife Service, Edmonton,
571 Alberta, Canada.
- 572 Holroyd, P., and T. Simieritsch. 2009. The waters that bind us: transboundary implications of
573 oil sands development. The Pembina Institute, Drayton Valley, Alberta, Canada.
- 574 Hunter, M. L., Jr. 1990. Wildlife, forests, and forestry: principles of managing forests for
575 biological diversity. Prentice-Hall, Englewood Cliffs, New Jersey.
- 576 Hunter, M. L., Jr. 2002. Fundamentals of conservation biology. Blackwell Science, Malden,
577 Massachusetts, USA.
- 578 James, A. R. C., S. Boutin, D. M. Hebert, and A. B. Rippin. 2004. Spatial separation of caribou
579 from moose and its relation to predation by wolves. *Journal of Wildlife Management* 68:
580 799–809.
- 581 Kelly, G. 2009. The oil sands—Canada's path to clean energy? Kingsley Publishing, Calgary,
582 Alberta, Canada.
- 583 Lagimodiere, M., and B. Eaton. 2009. Fish and fish habitat indicators for the Lower Athabasca
584 Regional Plan (LARP): description, rationale and modelling coefficients. Prepared for
585 the Lower Athabasca Regional Plan by the Fish Element Team, Government of Alberta.

- 586 Lee, P., and R. Cheng. 2009. Bitumen and biocarbon: land use changes and loss of biological
587 carbon due to bitumen operations in the boreal forests of Alberta, Canada. *Global Forest*
588 *Watch Canada*, Edmonton, Alberta, Canada.
- 589 Moss, E. H. 1983. *Flora of Alberta: a manual of flowering plants, conifers, ferns and fern allies*
590 *found growing without cultivation in the Province of Alberta, Canada. Second edition.*
591 *Revised by J. G. Packer. University of Toronto Press, Toronto, Ontario, Canada.*
- 592 Naill, R. 1973. The discovery cycle of a finite resource: a case study of U.S. natural gas. Pages
593 213-256 *in* D. L. Meadows and D. H. Meadows, editors. *Toward global equilibrium:*
594 *collected papers. Wright-Allen Press, Cambridge, Massachusetts, USA.*
- 595 Natural Regions Committee. 2006. *Natural regions and subregions of Alberta. Compiled by D.*
596 *J. Downing and W. W. Pettapiece. Government of Alberta, Canada. Publication No.*
597 *T/852.*
- 598 Nelson, J. S., and M. J. Paetz. 1992. *The fishes of Alberta. Second edition. University of*
599 *Alberta Press, Edmonton, Alberta, Canada.*
- 600 Nikiforuk, A. 2008. *Tar sands: dirty oil and the future of a continent. Greystone Books,*
601 *Vancouver, British Columbia, Canada.*
- 602 Noss, R. F., and A. Y. Cooperrider. 1994. *Saving nature's legacy: protecting and restoring*
603 *biodiversity. Island Press, Washington, D.C., USA.*
- 604 Park, D., M. Sullivan, E. M. Bayne, and G. Scrimgeour. 2008. Landscape-level stream
605 fragmentation caused by hanging culverts along roads in Alberta's boreal forest.
606 *Canadian Journal of Forest Research* 38:1189–1197.
- 607 Pattie, D. L., and R. S. Hoffmann. 1992. *Mammals of the North American parks and prairies.*
608 *Second edition. Self-published.*

- 609 Possingham, H. P., I. R. Ball, and S. Andelman. 2000. Mathematical methods for identifying
610 representative reserve networks. Pages 291–305 *in* S. Ferson and M. Burgman, editors.
611 Quantitative methods for conservation biology. Springer-Verlag, New York, New York,
612 USA.
- 613 Post, J. R., M. Sullivan, S. Cox, N. P. Lester, C. J. Walters, E. A. Parkinson, A. J. Paul, L.
614 Jackson, and B. J. Shuter. 2002. Canada's recreational fisheries: the invisible collapse?
615 Fisheries Management 27:6–17.
- 616 Prebble, P., A. Coxworth, T. Simieritsch, S. Dyer, M. Huot, and H. Walsh. 2009. Carbon copy:
617 preventing oil sands fever in Saskatchewan. Pembina Foundation, Saskatchewan
618 Environmental Society, and Canadian Parks and Wilderness Society.
- 619 Raynolds, M., M. McCulloch, and R. Wong. 2006. Carbon neutral by 2020: a leadership
620 opportunity in Canada's oil sands. Pembina Institute, Drayton Valley, Alberta, Canada.
- 621 Russell, A.P., and A. M. Bauer. 1993. The amphibians and reptiles of Alberta. University of
622 Calgary Press, Calgary, Alberta, Canada.
- 623 Schindler, D. W., K. H. Mills, D. F. Malley, D. L. Findlay, J. A. Shearer, I. J. Davies, M.
624 A. Turner, G. A. Lindsey, and D. R. Cruikshank. 1985. Long-term ecosystem stress: the
625 effects of years of experimental acidification of a small lake. Science 228:1395–1401.
- 626 Schindler, D. W., W. F. Donahue, and J. P. Thompson. 2007. Future water flows and human
627 withdrawals in the Athabasca River. Pages 1–38 *in* Running out of steam? Oil sands
628 development and water use in the Athabasca River-watershed: science and market based
629 solutions. University of Alberta Environmental Research and Studies Centre, Edmonton,
630 Alberta, Canada.

- 631 Schmiegelow, F. K. A., S. G. Cumming, S. Harrison, S. Leroux, K. Lisgo, R. Noss, and B.
632 Olsen. 2006. Conservation beyond crisis management: a reverse-matrix model.
633 BEACONs Discussion Paper No. 1, 2006.
634 <<http://www.beaconsproject.ca/resources.htm>>. Accessed 13 Oct 2009.
- 635 Schneider, R., and S. Dyer. 2006. Death by a thousand cuts: impacts of in situ development on
636 Alberta's boreal forest. Pembina Institute and Canadian Parks and Wilderness Society.
- 637 Semenchuk, G. P. 1992. The atlas of breeding birds of Alberta. Federation of Alberta
638 Naturalists, Edmonton, Alberta, Canada.
- 639 Smith, H. C. 1993. Alberta mammals: an atlas and guide. The Provincial Museum of Alberta,
640 Edmonton, Alberta, Canada.
- 641 Sullivan, M. G. 2003. Active management of walleye fisheries in Alberta: dilemmas of
642 managing recovering fisheries. North American Journal of Fisheries Management
643 23:1343–1358.
- 644 Sullivan, M. G. 2009. Conceptual framework for public motorized access management
645 strategies for ALCES-LARP modelling project. Unpublished report. Alberta Sustainable
646 Resource Development.
- 647 Sustainable Ecosystem Working Group [SEWG]. 2008. Terrestrial ecosystem management
648 framework for the regional municipality of Wood Buffalo. Cumulative Environmental
649 Management Association. <<http://www.cemaonline.ca/content/view/75/182/>>. Accessed
650 13 Oct 2009.
- 651 Trew, D. O. 1995. The sensitivity of Alberta lakes to acidic deposition. Proceedings of the
652 Fourth Annual Workshop of the Alberta Lake Management Society, 30 September–1
653 October 1995, Lac la Biche, Alberta, Canada.

- 654 Vitt, D. H., J. E. March, and R. B. Bovey. 1988. Mosses, lichens and ferns of northwest North
655 America. Lone Pine Press, Edmonton, Alberta, Canada.
- 656 Wells, J., S. Casey-Lefkowitz, G. Chavarria, and S. Dyer. 2008. Danger in the nursery: impact
657 on birds of tar sands oil development in Canada's boreal forest. Natural Resources
658 Defense Council, Boreal Songbird Initiative, and Pembina Institute.
- 659

660 **Tables**

661 Table 8-1. Best practices included in land use simulations.

Best practice	Description (values in parentheses represent BAU, realistic BP, and optimistic BP, respectively)
Minimize energy sector footprint	<p data-bbox="656 495 1409 600">Greater dependency on directional drilling, in order words, placing more wells on a single pad (10, 17, 25 wells/pad)</p> <p data-bbox="656 642 1409 821">Develop pipelines along road corridors to reduce overall footprint (20%, 40%, and 60% of pipeline footprint overlapping with roads)</p> <p data-bbox="656 863 1409 968">Periodic reclamation of old seismic lines (0%, 5%, and 10% of existing seismic lines every 5 years)</p> <p data-bbox="656 1010 1409 1115">Minimize the width of new seismic lines to accelerate reclamation time (21, 10, and 2 year lifespan)</p> <p data-bbox="656 1157 1409 1262">Accelerated reclamation of well pad after production has ceased (10, 5, and 0 year reclamation lag)</p> <p data-bbox="656 1304 1409 1402">Accelerate reclamation of surface mines after production has ceased (10, 5, and 0 year reclamation lag)</p>
Minimize forestry sector footprint	<p data-bbox="656 1444 1409 1549">Accelerate in block road reclamation after timber harvest (10, 7.5, and 5 years reclamation lag)</p> <p data-bbox="656 1591 1409 1621">Larger cut blocks to reduce road requirements</p>

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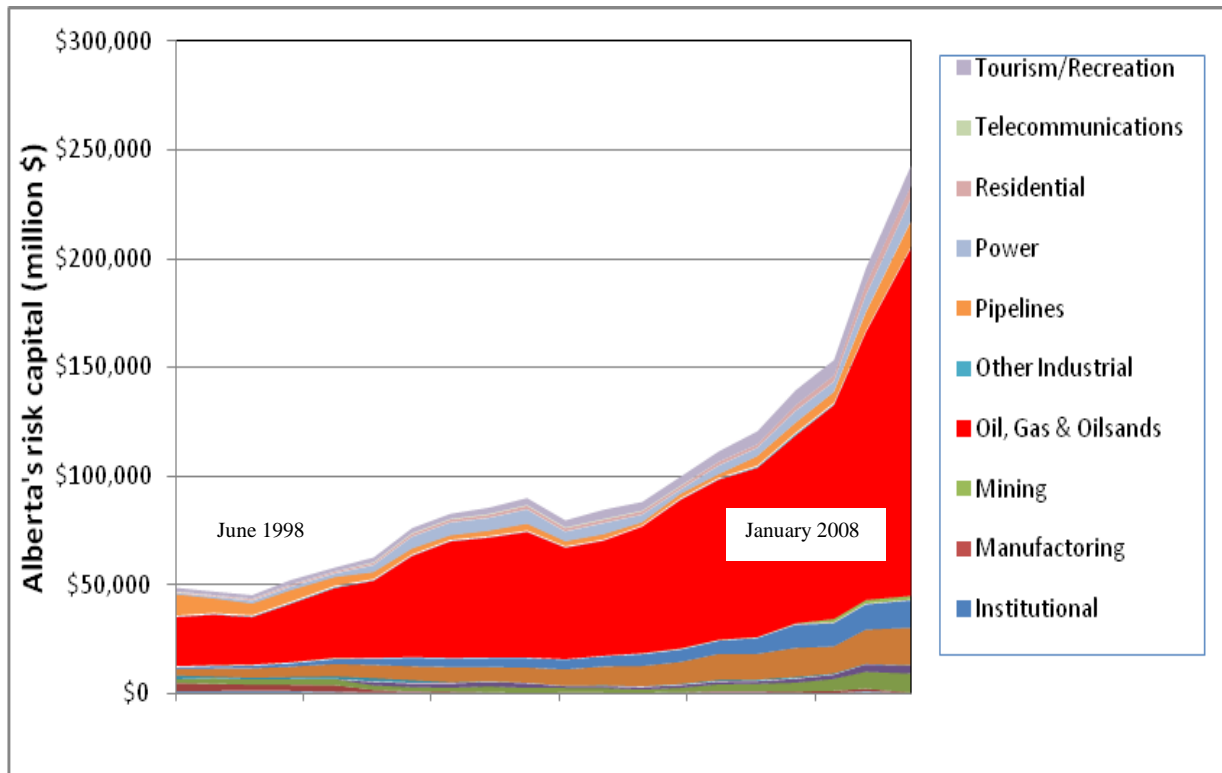
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Table 8-1. Continued.

Best practice	Description (values in parentheses represent BAU, realistic BP, and optimistic BP, respectively)
Maintain older forest	Avoid harvesting some older stands to maintain older forest (older forest targets of 0, 9%, and 18% of merchantable forest). Attempt to offset timber supply cost of old forest protection through in genetic improvement and increased utilization standards (0, 5%, and 10% increase in merchantable timber volume)
Hung culvert replacement	Remove and replace hung culverts (2, 5, and 10% annual replacement of hung culverts)
Emissions reduction	<p>Reduce CO₂ emissions by the mineable bitumen sector (0.55, 0.40, and 0.25 tonne/m³ bitumen production)</p> <p>Reduce CO₂ emissions by the in-situ bitumen sector (0.535, 0.405, and 0.275 tonne/m³ bitumen production)</p> <p>Reduce NO_x emissions associated with bitumen production (0.000693, 0.000589, 0.000485 tonne/m³ bitumen production)</p>
Water conservation	<p>Reduce net water consumption by the mineable bitumen sector (2.5, 2.15, 1.75 m³/m³ bitumen production)</p> <p>Reduce net water consumption by the in-situ bitumen sector (7.4, 5.0, 2.7 m³/m³ bitumen production)</p>

664 **Figures**

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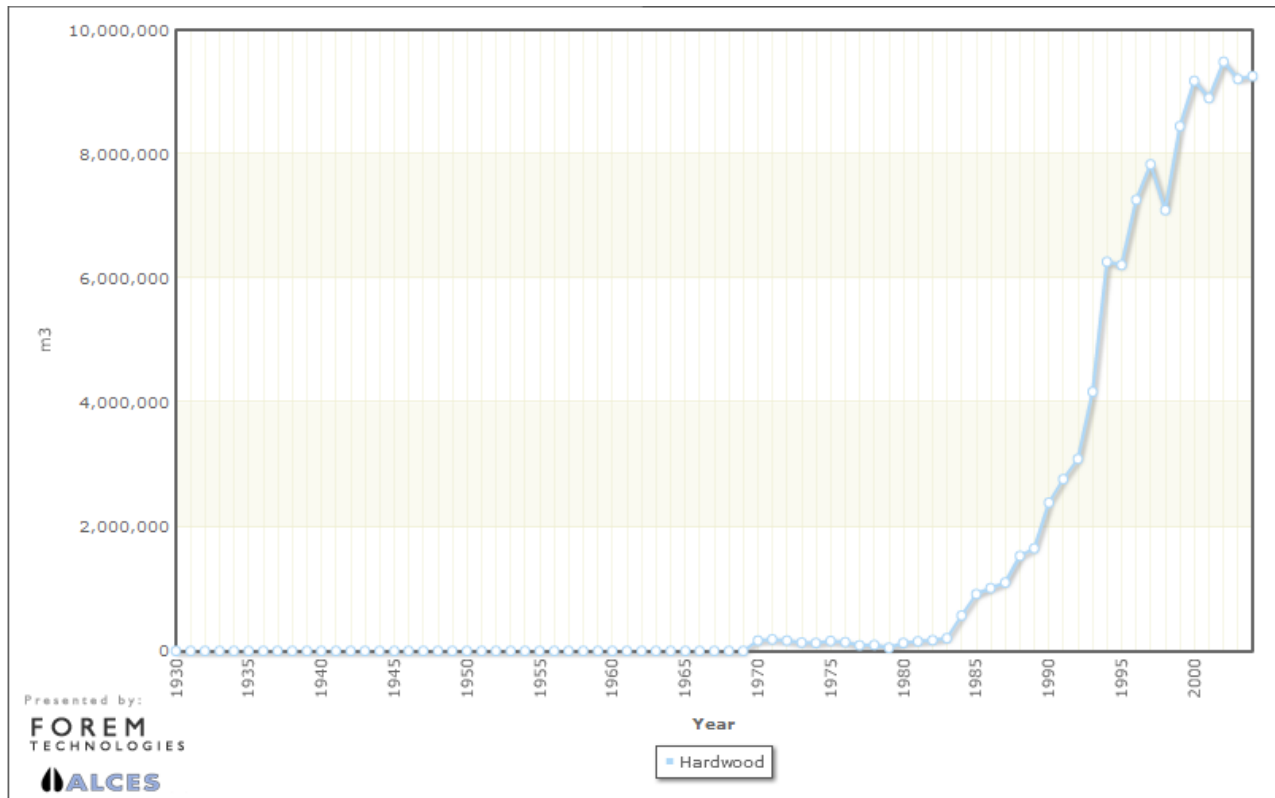


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667 Figure 8-1. Alberta's risk capital by sector between 1998 and 2008.

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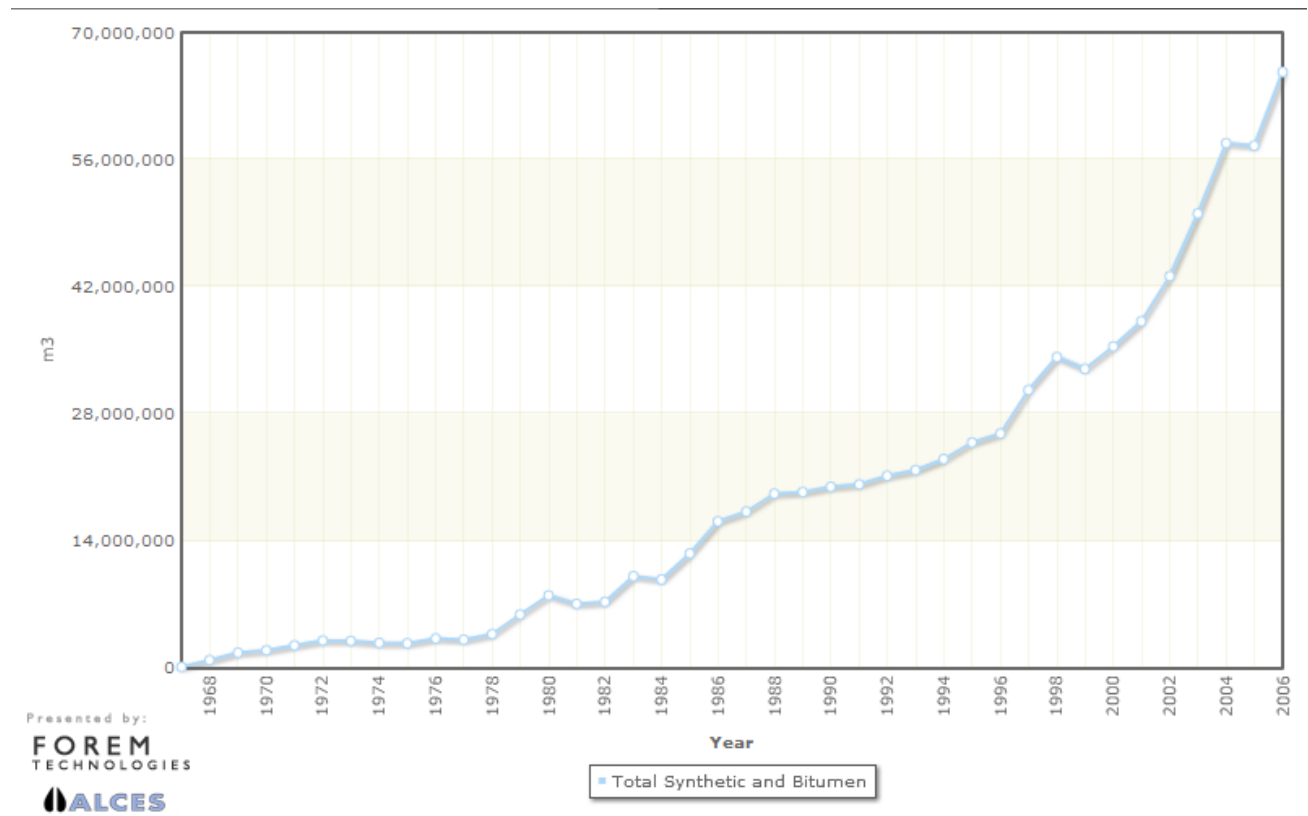


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671 Figure 8-2. Temporal trend in annual hardwood harvest in Alberta. The annual allowable
672 harvest volume of Al-Pac in northeast Alberta represents the single largest allocation in Alberta
673 at 3.5 M m³/year.

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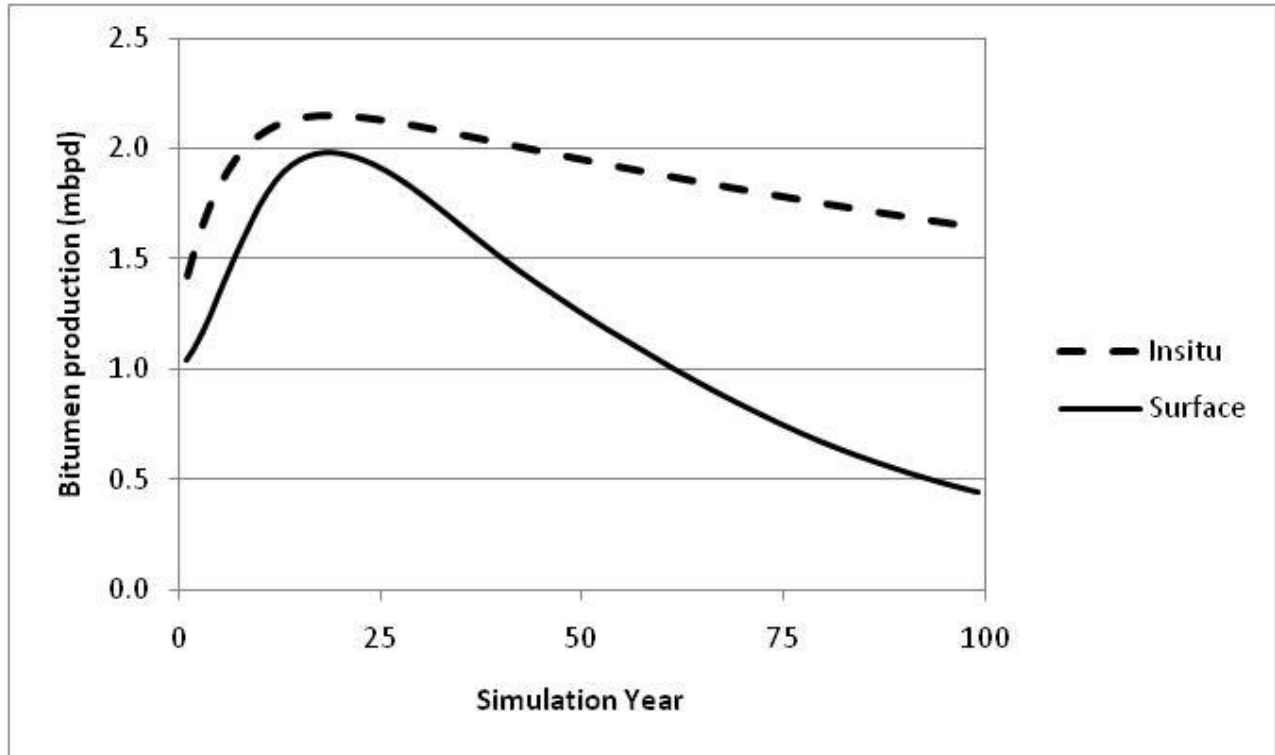


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677 Figure 8-3. Temporal trend in bitumen (synthetic and raw) in Alberta. The majority of bitumen
678 is produced in northeast Alberta from the Athabasca and Cold Lake deposits.

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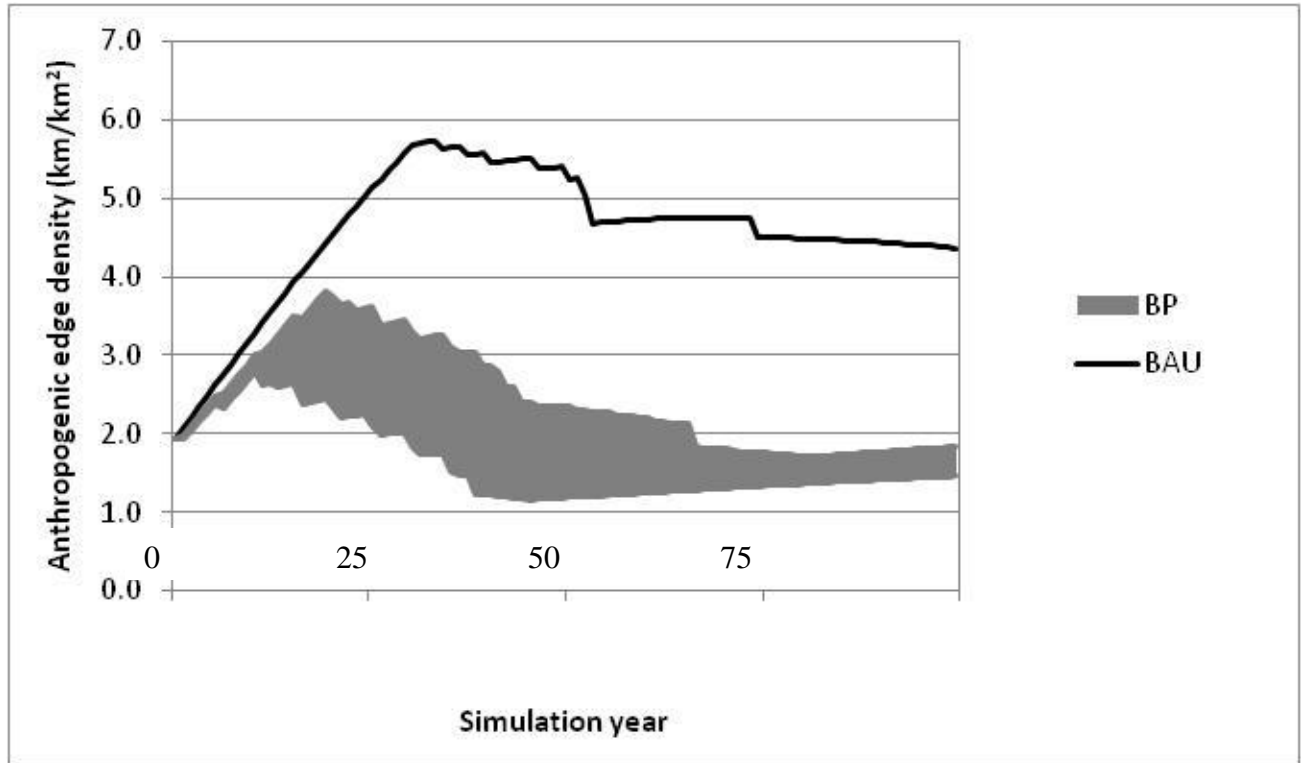
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682 Figure 8-4. Projected annual in-situ and mineable bitumen production from the study area over
683 the next 100 years.

684



685

686 Figure 8-5. Simulated anthropogenic edge density in the study area over the next 100 years

687 under business as usual practices (BAU) and best practices (BP) land-use scenarios in ALCES.

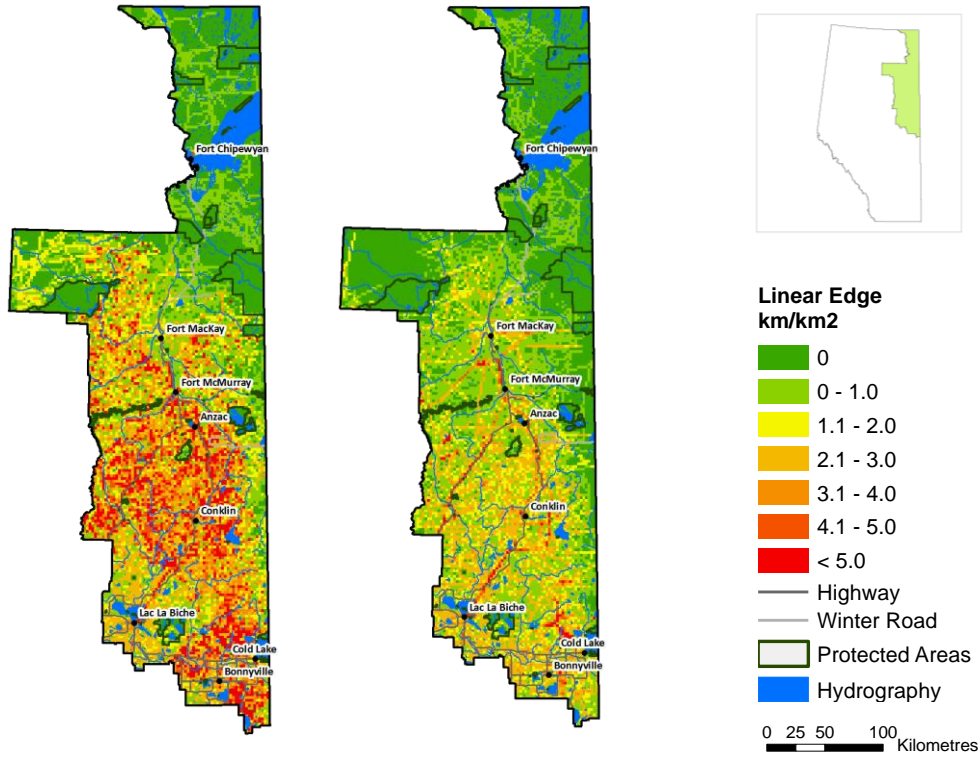
688 Best practices are presented as a band to reflect the range in indicator response associated with

689 realistic to optimistic implementation of the best practices.

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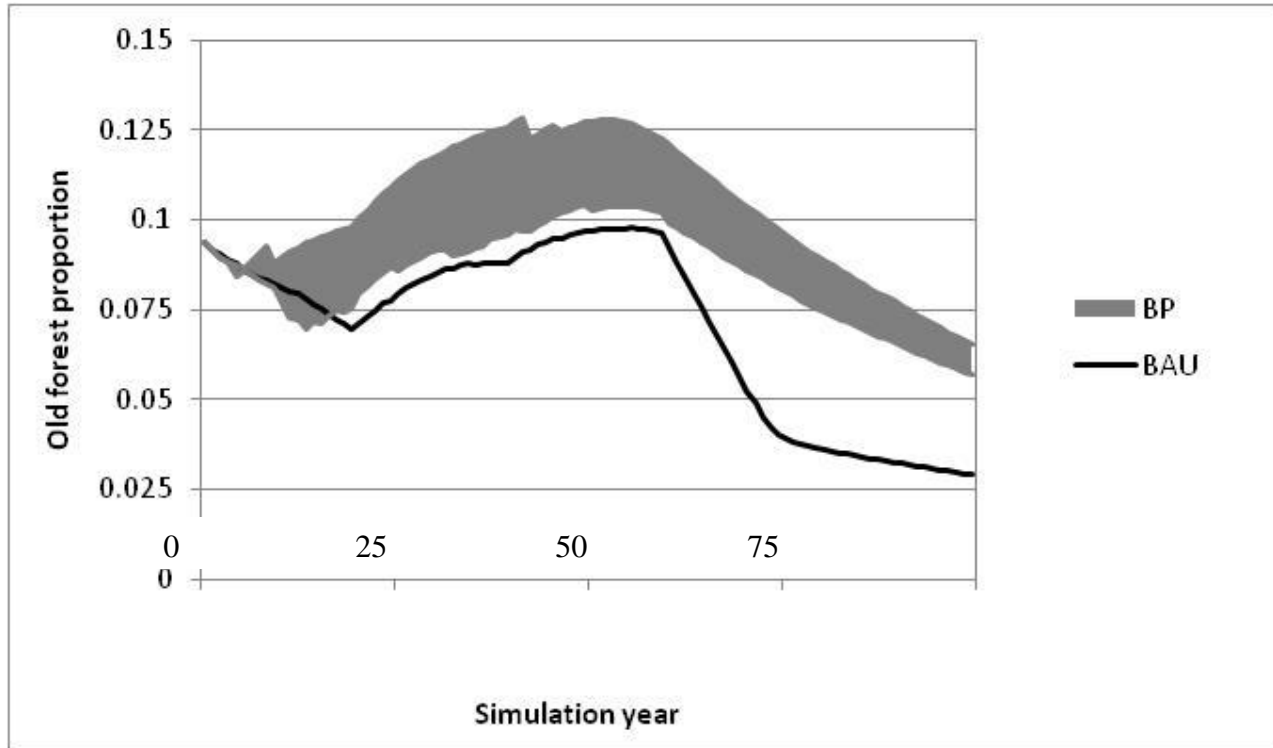
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703 Figure 8-6. Anthropogenic edge density at simulation year 60 under business as usual (left) and

704 high best practices (right) scenarios.

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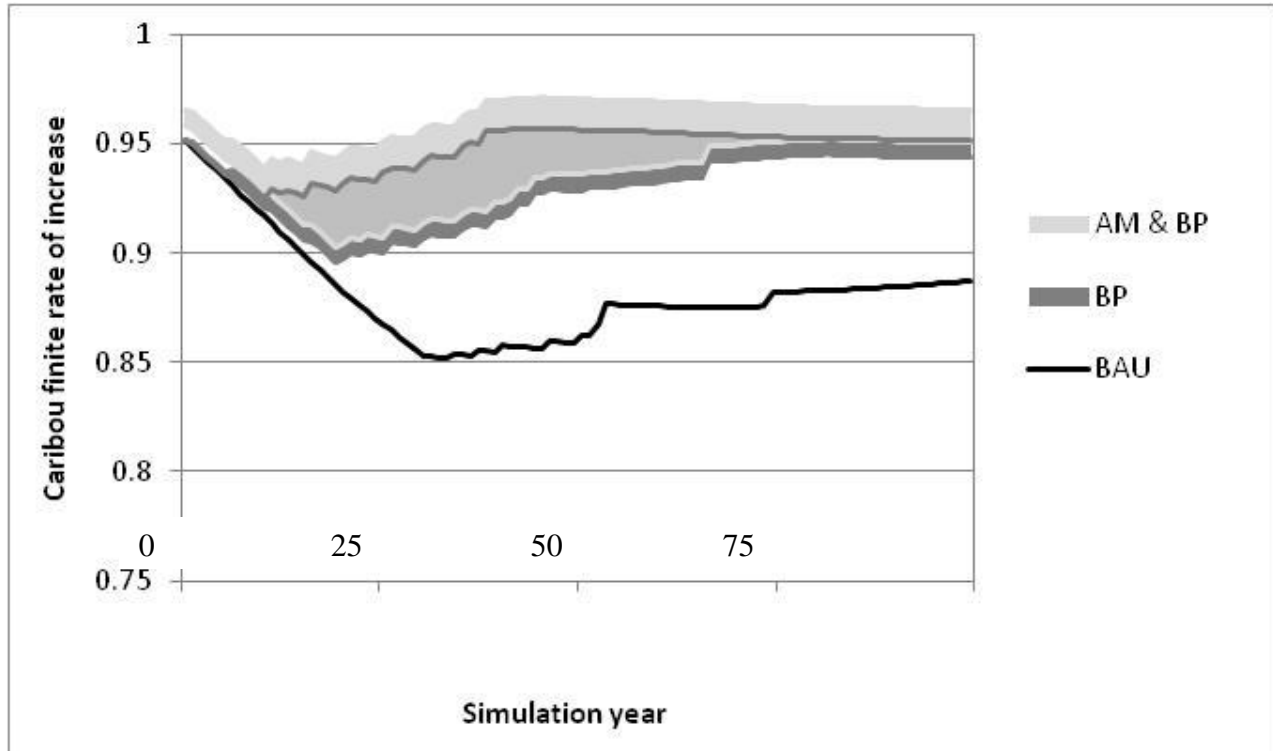


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708 Figure 8-7. Simulated proportion of forest that is older than 120 years in the study area over the
709 next 100 years under business as usual practices (BAU) and best practices (BP) land-use
710 scenarios in ALCES. Best practices are presented as a band to reflect the range in indicator
711 response associated with realistic to optimistic implementation of the best practices.

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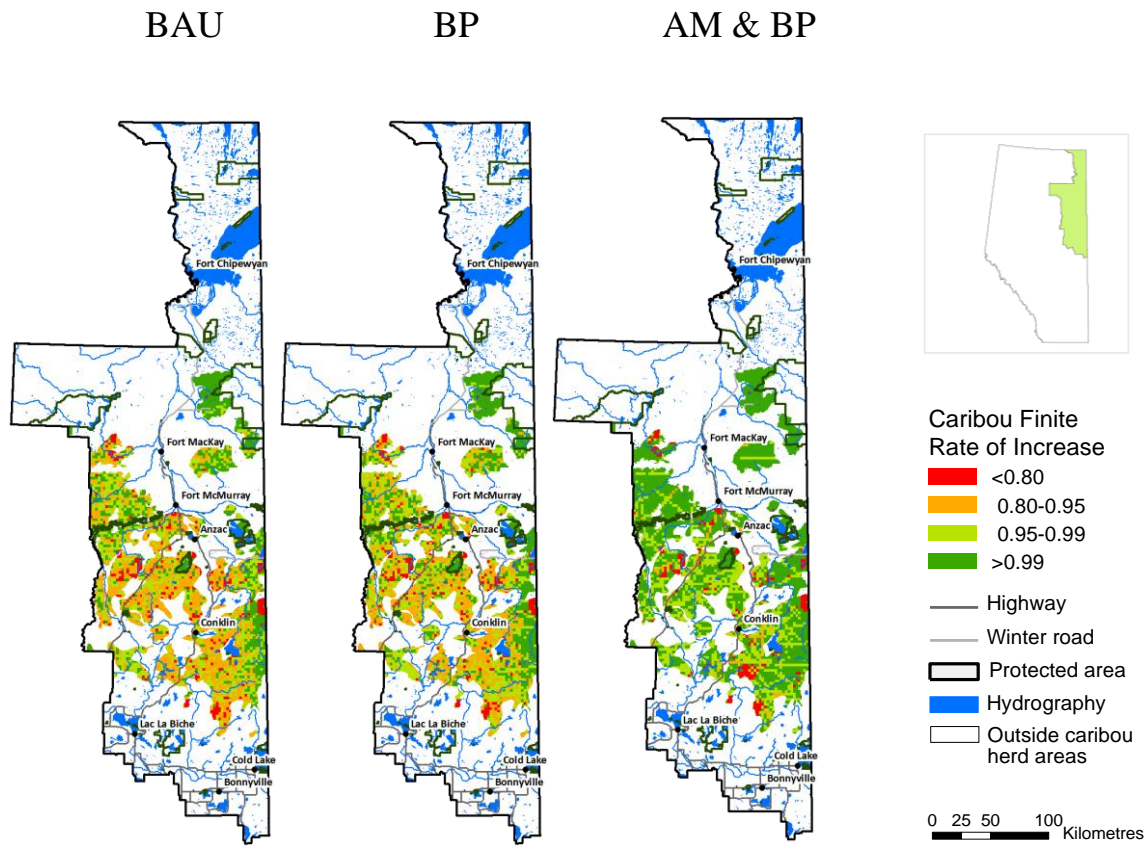
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715 Figure 8-8. Simulated caribou finite rate of increase area over the next 100 years under business
716 as usual practices (BAU), best practices (BP), and access management plus best practices (AM &
717 BP) land-use scenarios in ALCES. BP and AM & BP are presented as bands to reflect the range
718 in indicator response associated with realistic to optimistic implementation of the mitigation
719 strategies.

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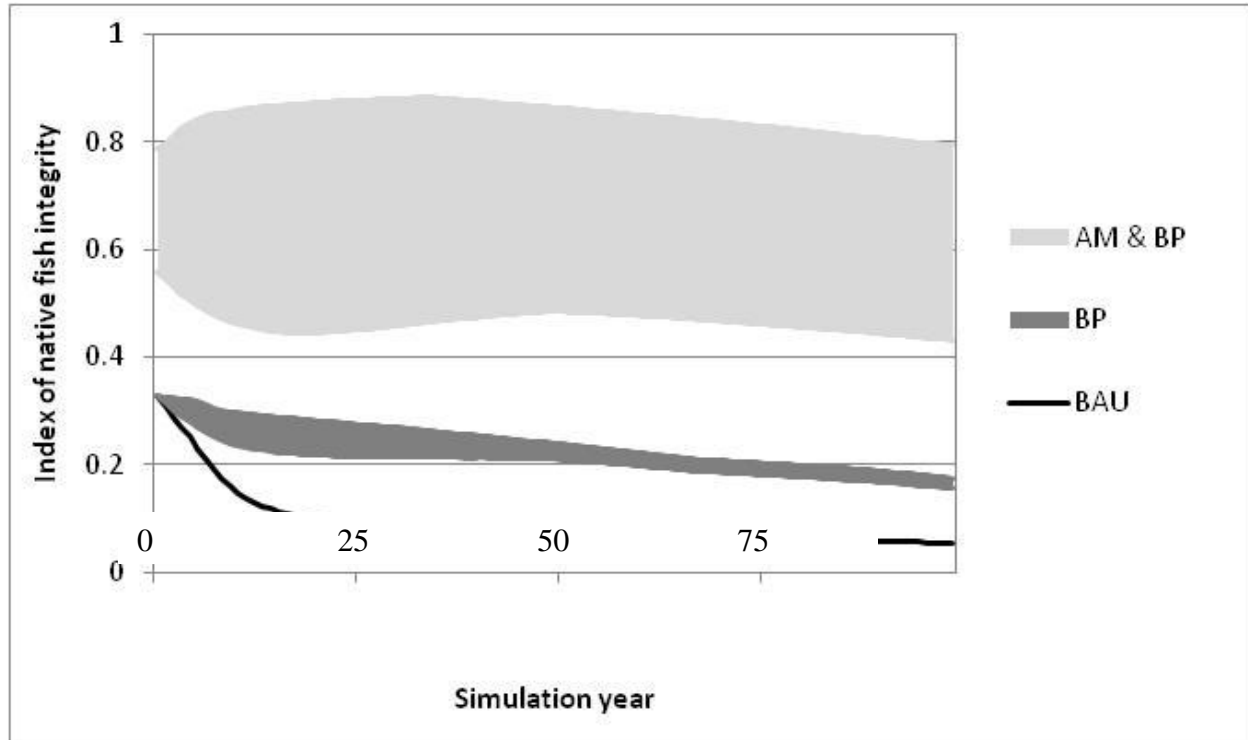
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Figure 8-9. Caribou finite rate of increase across the study area at simulation year 60 under business as usual (BAU), high best practices (BP), and high access management and high best practices (AM & BP) scenarios. Herd areas are from the University of Alberta ILM Lab (2008).

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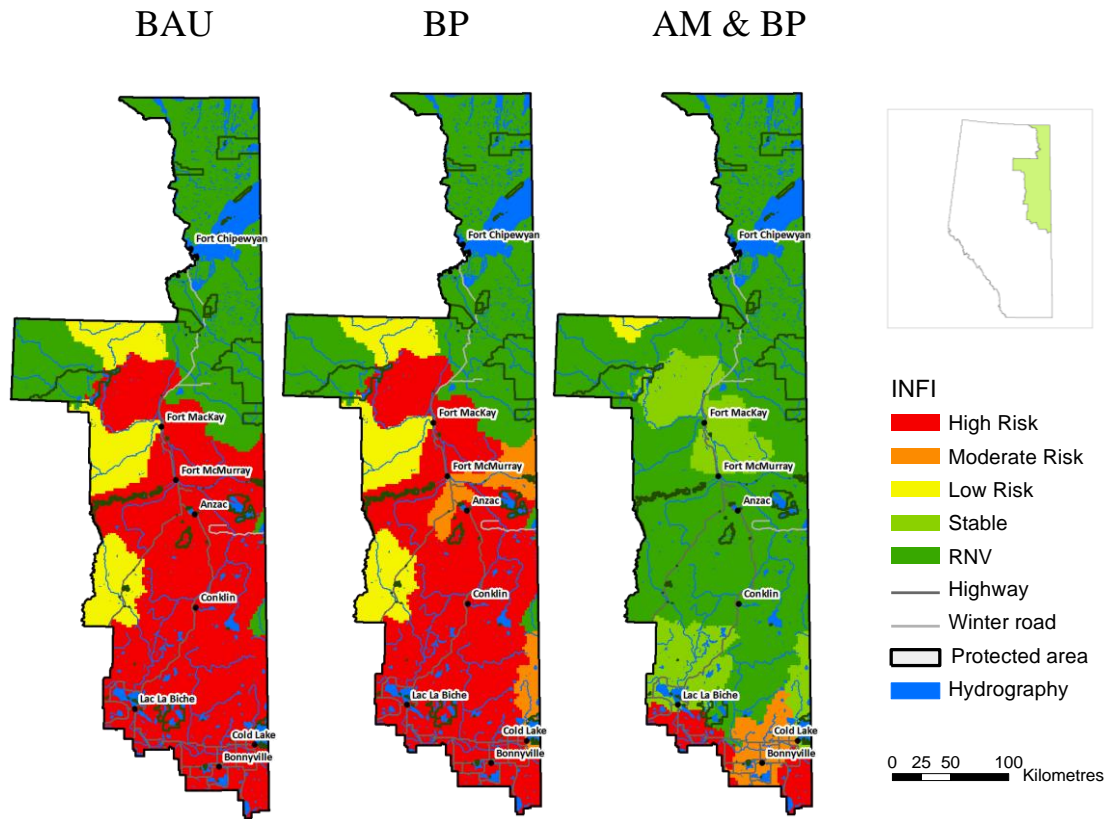
731 Figure 8-10. Simulated index of native fish integrity over the next 100 years under business as
732 usual practices (BAU), best practices (BP), and access management plus best practices (AM &
733 BP) land-use scenarios in ALCES. BP and AM & BP are presented as bands to reflect the range
734 in indicator response associated with realistic to optimistic implementation of the mitigation
735 strategies.

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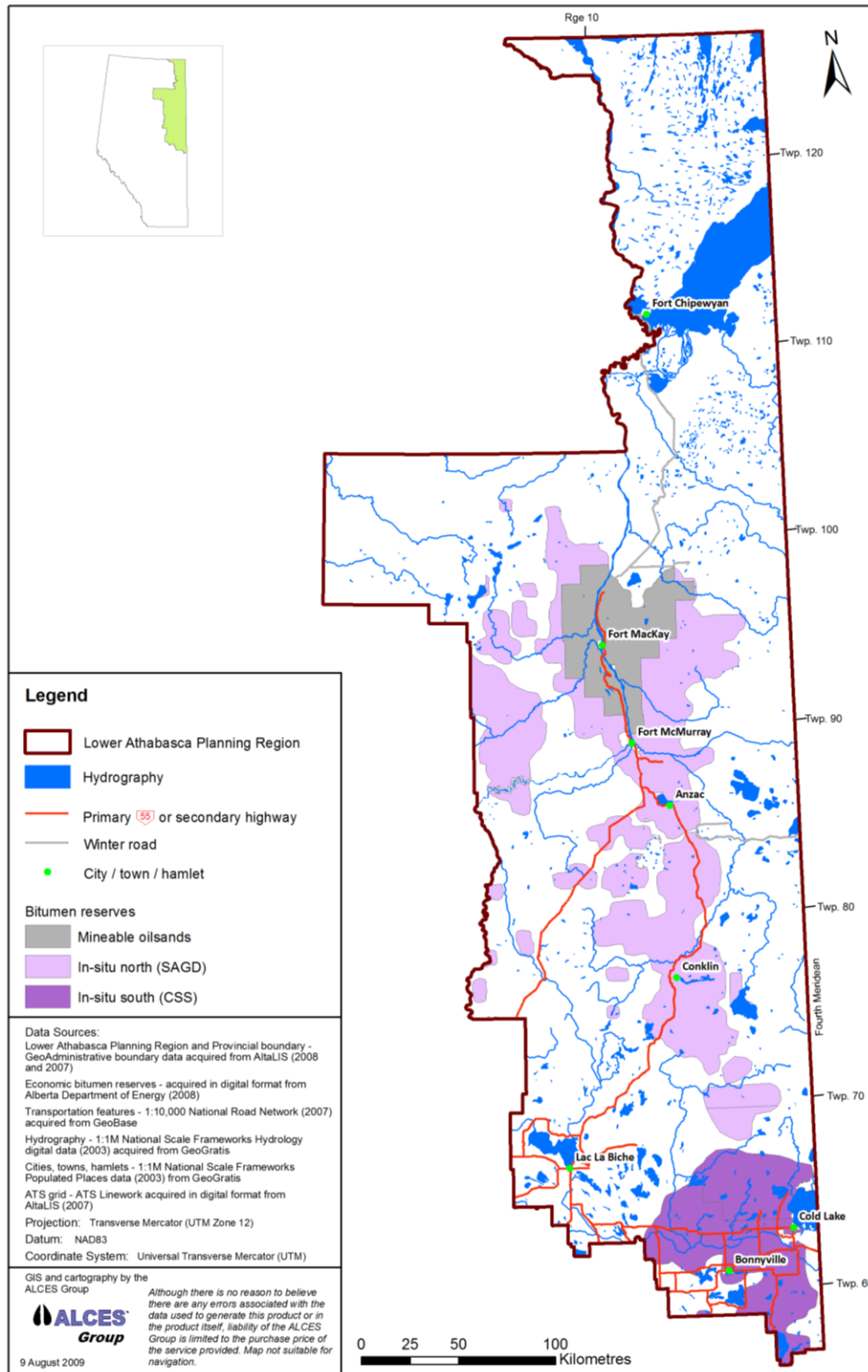
741 Figure 8-11. Status of the index of native fish integrity (INFI) across the study area at simulation

742 year 60 under business as usual (BAU), high best practices (BP), and high access management

743 and high best practices (AM & BP) scenarios.

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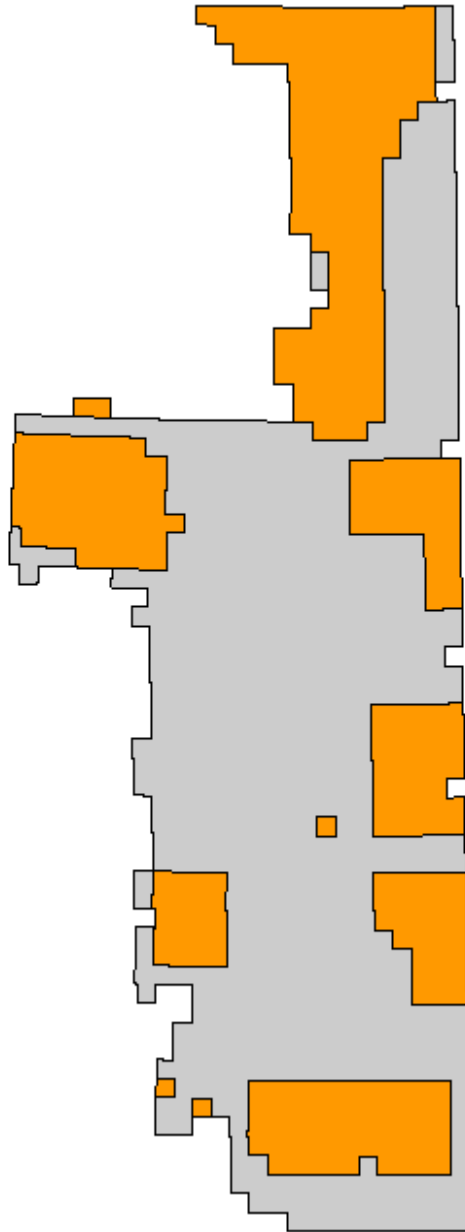
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747 Figure 8-12. In-situ and mineable oilsands deposits in the study area.

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750 Figure 8-13. Least cost conservation area network (orange) identified using MARXAN for

751 representing 30% of each natural subregion and land cover type.